Combined photolysis and catalytic ozonation of dimethyl phthalate in a high-gravity rotating packed bed

Chia-Chi Chang\textsuperscript{a}, Chun-Yu Chiu\textsuperscript{b}, Ching-Yuan Chang\textsuperscript{a,}\textsuperscript{*,} Chiung-Fen Chang\textsuperscript{c}, Yi-Hung Chen\textsuperscript{d}, Dar-Ren Ji\textsuperscript{a}, Yue-Hwa Yu\textsuperscript{a}, Pen-Chi Chiang\textsuperscript{a}

\textsuperscript{a} Graduate Institute of Environmental Engineering, National Taiwan University, Taipei 106, Taiwan
\textsuperscript{b} Department of Cosmetic Science and Application, Lan-Yang Institute of Technology, I-Lan 261, Taiwan
\textsuperscript{c} Department of Environmental Science and Engineering, Tunghai University, Taichung 407, Taiwan
\textsuperscript{d} Department of Chemical and Material Engineering, National Kaohsiung University of Applied Science, Kaohsiung City 807, Taiwan

1. Introduction

Endocrine disrupting chemicals (EDCs, also called environmental hormones) have been spread over the environment and influenced the generative function of some species on the earth. Most of EDCs are persistent organic pollutants (POPs) which exhibit chronic toxicity, bioaccumulation and are difficult to be decomposed by organisms in the environment. Among the EDCs, phthalate acid esters (PAEs) are chemical compounds most widely used in the plastic industrial production. The major applications of PAEs are for the use as plasticizers for polyvinyl chloride (PVC) resins, adhesives and cellulose film coating (counting about 85% of the whole production). The short-chained ester such as dimethyl phthalate (DMP) is among the most frequently identified PAEs in diversified environmental samples including upper layer of marine water, surface waters and sediments. Esters are found to accumulate in the environment and to be toxic to a variety of aquatic organisms, which are at the base of the food chain in both marine and surface water environments. Destruction of these compounds requires the application of processes having high decomposition abilities such as those being able to provide highly reactive hydroxyl radicals. Recently, researches showed that phthalate esters in secondary effluent and surface water can be decomposed via O$_3$/UV processes \cite{1,2}.

Ozonation processes have been widely applied to water and wastewater treatment. In aqueous solution, ozone acts on various compounds by direct ozonation and radical reaction. Molecular ozone is a powerful oxidant agent which reacts selectively with organic compounds in water. The organic compounds with molecular structure consisting of multiple bonds (for example C=\textsubscript{2}C, C=C\textsubscript{2}=O, R and \textsubscript{2}C=C\textsubscript{2}=X) or atom carrying a negative charge (such as N, P, O, S and nucleophilic carbon) are highly reactive with ozone \cite{3}. The ozonation system should be able to form highly reactive radical species from the decomposition of molecular ozone in water so as to destroy the POPs via free radical chain reaction. Alternative ozonation processes using catalysts of transition metals also have been applied for the degradation of organics. Catalytic ozonation can be classified according to (1) activation of ozone via metal ions in solution and (2) heterogeneous catalytic ozonation in the presence of metal oxides or metals on supports \cite{4}. The use of metal
ions as the homogeneous catalyst has no mass transfer resistance between the reactants and catalyst. However, the metal ion remains in the solution and is not easy to be recovered or removed. Solid catalysts which are packed in the packed bed are able to be recycled and regenerated. The catalysts with metals such as Pt, Ru, Rh, Pd and Ag were reported to be effective for the decomposition of organic compounds in gaseous as well as aqueous-phase ozonation. The order of catalytic activity of metals supported on Al₂O₃ was found to be: Pt > Pd > Ag > Ru ~ Rh ~ Ir > Ni > Cd > Mn > Fe > Cu > Zn ~ Zr [5]. In this study, platinum-containing catalyst, Dash 220N, was chosen as the catalyst to produce the hydroxyl radicals from aqueous ozone.

Ozonation is limited by its selective reactivity and low gas–liquid mass transfer rate. The high–gravity rotating packed bed (HGRPB, HG or Higee) was used in this study as a gas–liquid contactor to enhance the mass transfer rate between the phases. The values of volumetric gas–liquid mass transfer coefficients achievable in HGRPB are one or two orders of magnitude higher than those in conventional packed beds [6]. Recent studies [7–9] showed that HGRPB has some advantageous characteristics such as (1) fast renewable rate on the surface of packed materials, (2) high gas–liquid mass transfer coefficients, (3) short time to reach steady state, (4) low overflow rate, (5) short hydraulic retention time (HRT) and (6) thin liquid film.

HGRPB is also applicable to absorption, adsorption, desorption, distillation, polymer devolatilization, bio-oxidation, reactive crystallization, stripping, extraction and other separation processes [10]. Recently, HGRPB also has been used for micro-mixing and emulsion [11,12].

Previous studies point out that HGRPB has a higher mass transfer rate in gas–liquid system than conventional gas–liquid contactor. Thus, the ozonation process can be combined with HGRPB to enhance the ozone mass transfer from gas phase to liquid phase. Lin and Liu [13] found that the centrifugal force can facilitate the decolorization efficiency of dyestuff Reactive Blue 19 (RB19) because the mass transfer rate of ozone is enhanced via increasing the interfacial area and decreasing the resistance of ozone transfer. Chen et al. [14] used RBP as well as completely stirred tank reactor (CSTR) as ozone gas–liquid contactor to treat CI Reactive Black 5 (RB5). The results indicated that the mass transfer coefficient of RBP (k_La = 0.0975 s⁻¹) is significantly higher than that of CSTR (k_La = 0.0248 s⁻¹).

In this research, the HGRPB was applied to the catalytic ozonation processes using Pt/γ-Al₂O₃ to decompose the organic compound of DMP in aqueous solution. System performances for the degradation of DMP via various advanced oxidation processes (AOPs) using HGRPB were examined and compared. These HGRPB assisted AOPs include HGRPB-ozonation (HG-OZ), Pt catalytic HG-OZ (HG-Pt-OZ), ultra violet HG-OZ (HG-UV-OZ) and HG-UV-Pt-OZ. The results can provide insight information for the feasibility evaluation of applying a HGRPB to the said AOPs.

2. Experimental

2.1. Chemicals

DMP with chemical formula of C₁₀H₁₀O₄, molecular weight of 194.19 and purity >99% was purchased from Hayashi Pure Chemical Industries Ltd. (Osaka, Japan). The water solubility of DMP is 4000 mg L⁻¹ at 25 °C (1 atm). The DMP was subjected to DI water prior to the use with the concentration (C_{DMP}) of 100 mg L⁻¹. The initial pH of the DMP solution is about 7.5. The potassium iodide (KI), sodium chloride (NaCl) and acetonitrile (CH₃CN) used were of analytical grade (J.T. Baker, Phillipsburg, NJ, USA). Ozone was decomposed to form radicals via a commercial Pt/γ-Al₂O₃ catalyst Dash 220N (N.E. Chemical Co., Japan), which is a platinum-containing catalyst. Characteristic of the Dash 220N is shown in Table 1 [15]. Dash 220N is a spherical catalyst 3–4 mm in diameter. The bulk density of the Dash 220N is 0.77 g cm⁻³. The content of platinum in the alumina sphere is about 0.23 wt.%.

2.2. Instrumentation

The HGRPB system (Fig. 1) was carried out via semi-batch operation. The system has two major parts, namely RPB and photoreactor. The RPB consists of a packing chamber, a rotor and a stationary housing. The DMP solution flows outward from the inner edge of
the packing chamber subjected to the centrifugal force. Gaseous ozone flows inward countercurrently from the outer edge of the packed bed by the pressure-driving force. The experiments were performed with the volume ($V_L$) of DMP solution of 1 L and total reaction time ($t$) of 60 min. To identify the catalytic ability of platinum in ozonation system, two packing materials of Dash 220N and spherical glass beads were used and tested in the packed bed. The diameter of glass beads is about the same as that of Dash 220N. The RPB is 2 cm in height and 5.9 cm in diameter. In the experiments, the RPB was operated with rotating speed ($N_r$) from 600 to 1800 rpm, which provided gravitational force of 23.76–213.91 g (232.92–2096.31 m s$^{-2}$). The DMP solution was fed to RPB by a peristaltic pump at 462 mL min$^{-1}$ recycle rate ($Q_{LR}$) and contacted with gaseous ozone and catalysts in the bed.

The photoreactor is made of Pyrex glass with the dimensions of diameter and height of 5 and 30 cm, respectively. A UV lamp (TUV-16W, Phillips, Tokyo, Japan) was placed at the center of the UV reactor and shielded by a quartz jacket. The UV irradiation was introduced to enhance the decomposition of ozone to form the OH$^\cdot$ radical. All the experiments were controlled at 25°C with water jacket around the reactor. The DMP solution was pumped and recycled continuously through a closed loop connected with all sensors of pH, oxidation reduction potential (ORP) and dissolved ozone (DO$_3$) monitors. Liquid samples were taken from a sampling port for chemical analyses.

The ozone was generated from dried oxygen with purity of 99.99% via corona discharge using an ozone generator (Model LAB2B, Ozonia, Duebendorf, Switzerland). The ozone inlet flow rate ($Q_G$) was controlled via a mass flow controller (MFC, Model 5850E, Brooks, Hatfield, PA, USA) at 1 L min$^{-1}$. The inlet and outlet gaseous ozone concentrations were measured via a spectrophotometer (UV mini-1240, Shimadzu, Kyoto, Japan) with the absorbance of ozone measured in a 2 mm flow-through quartz cell at the wavelength 258 nm. An extinction coefficient of 3000 M$^{-1}$ s$^{-1}$ was used to convert absorbances into concentration units [16]. The inlet ozone concentration ($C_{AG,\text{in}}$) was controlled at 25 mg L$^{-1}$. Ozone concentration in water was determined by DO$_3$ meter (Model 3600, Orbisphere, Traasadingen, Switzerland). The UV–vis irradiation intensity was measured by diffraction grating spectrometer (Model EPP 2000, StellarNet, Oldsmar, FL, USA).

The analyses of DMP solutions were performed using the high performance liquid chromatography (HPLC, Model 500, Viscotek, Houston, TX, USA) with 250 mm × 4.6 mm C18 column (LC-18, Supelco, Bellefonte, PA, USA) to separate the DMP and by-products of ozonation. The wavelength of UV/visible detector (Model 1706, Bio-Rad, Hercules, CA, USA) was set at 220 nm. Effluent is composed of CH$_3$CN and DI water (CH$_3$CN/DI water = 50/50) with flow rate controlled at 1.0 mL min$^{-1}$. Total organic carbon (TOC) was analyzed by TOC analyzer (Model 1010, O.I. analytical, College Station, TX, USA).

3. Results and discussion

3.1. Hydraulic test

Heterogeneous catalytic ozonation of DMP in HGRPB is a gas–liquid–solid reaction for which mass transfer and chemical reaction may play important roles. The mass transfer is dependent on the mixing of liquid, gaseous ozone and solid catalyst. The HGRPB is an efficient mixing apparatus with its high rotating speed. The HRT of HGRPB was determined by observing the variation of conductivity of tracer-containing effluent. The tracer (NaCl) was injected in the influent in the entrance of HGRPB. Fig. 2 shows the variations of HRT with $N_r$ and $Q_{LR}$. At low water feeding rate ($Q_{LR} = 188$ mL min$^{-1}$) and low rotating speed ($N_r = 300$ rpm), the HRT was about 10 s. The HRT reduced to 8 s as $N_r$ increased to 600 rpm for the case with $Q_{LR} = 188$ mL min$^{-1}$. At the same $Q_{LR}$, the HRTs were about the same for the cases with $N_r$ in the range of 600 to 1800 rpm. The HRTs were about 8, 6, 5, 4 and 4 at $Q_{LR}$ of 188, 463, 545, 750 and 857 mL min$^{-1}$. The result showed that at the same $N_r$, the HRT was strongly dependent on $Q_{LR}$. High $N_r$ leads to high centrifugal force assisting the fluid spraying outward.

Fig. 1. Schematic diagram of HGRPB system.
through the packed bed. Liquid volume in HGRPB is limited by the voidage of the packed bed ($\varepsilon_B$). The total volume of the packed bed is 213.88 cm$^3$ with the volume of packed catalyst of 72.05 cm$^3$ and stainless wire of 18.27 cm$^3$. This gives the free space of the packed bed of 123.56 cm$^3$. Thus, the $\varepsilon_B$ of HGRPB in this experiment was 57.77%. Theoretically, the HRTs without rotating would be 40, 16, 14, 10 and 9 s with $Q_{LR}$ of 188, 462, 545, 750 and 875 mL min$^{-1}$. Under high $N_r$, the liquid is speeded up via the centrifugal force, resulting in a significant decrease of HRT down to 4–10 s. However, the high $N_r$ of HGRPB can well mix the liquid with gaseous oxidant in a short time.

### 3.2. Photolysis and ozonation of DMP

The decomposition of DMP in HGRPB systems of HG, HG-UV, HG-OZ and HG-UV-OZ at $Q_G = 1\text{ L min}^{-1}$, $Q_{LR} = 100\text{ mL min}^{-1}$ and $N_r = 1200\text{ rpm}$ is shown in Fig. 3. In this part of experiment, the HGRPB was packed with glass spheres 4 mm in diameter. For the cases without ozone, nitrogen gas was introduced to maintain the same hydraulic conditions. DMP was not decomposed using HG and HG-UV. For the ozonation, the applied ozone dosage ($m_{A,in}$) and decomposition efficiency of DMP ($\eta_{DMP}$) are defined as follows:

$$m_{A,in} = Q_G \times C_{AG,in} \times t \times V_L^{-1}$$

$$\eta_{DMP} = \frac{(C_{DMP0} - C_{DMP})}{C_{DMP0}}$$

where $Q_G$, $C_{AG,in}$ and $V_L$ are gas flow rate, concentration of inlet ozone and volume of liquid sample, while $C_{DMP0}$ (or $C_0$) and $C_{DMP}$ (or $C$) are concentrations of DMP at time $t=0$ and $t$, respectively.

The values of $\eta_{DMP}$ for HG-OZ process reached about 80% in 30 min reaction time ($m_{A,in} = 600\text{ mg L}^{-1}$) and 96% in 60 min ($m_{A,in} = 1200\text{ mg L}^{-1}$), respectively, revealing effective decomposition of DMP via ozonation.

The emission spectrum of UV lamp (denoted as UV raw) for the photolysis experiments is shown in Fig. 4. Fig. 4a shows the absorption spectrum of DMP. Light source: 16 W UV lamp; $C_{DMP0} = 100\text{ mg L}^{-1}$. 

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**Fig. 2.** (a) Hydrolytic retention time (HRT) vs. rotating speed ($N_r$); (b) HRT vs. liquid flow rate ($Q_{LR}$).

**Fig. 3.** Time variations of $C/C_0$ for the decomposition of DMP in HGRPB systems. Initial concentration of DMP ($C_{DMP0}$ or $C_0$) = 100 mg L$^{-1}$, $Q_G = 1\text{ L min}^{-1}$, $Q_{LR} = 462\text{ mL min}^{-1}$, power of UV lamp = 16 W as noted in specification, $V_L = 1\text{ L}$, $N_r = 1200\text{ rpm}$, $C_{AG,in} = 20\text{ mg L}^{-1}$, $T = 25^\circ\text{C}$. ($\square$): HG only; ($\bigcirc$): HG-photolysis (HG-UV); ($\triangle$): HG ozonation (HG-OZ); ($\Delta$): HG-UV-OZ. HGRPB: high-gravity rotating packed bed.
raw emission spectrum of the UV lamp. It indicated that the low-pressure UV lamp employed is a multi-wavelengths light source with three regions of UV-A, UV-B and UV-C. The irradiation intensities of the UV lamp in UV-A (315–400 nm), UV-B (280–315 nm) and UV-C (200–280 nm) are 3.99, 1.59 and 3.73 W m⁻² counting the whole wavelength range of the respective regions. Fig. 4b and c are UV spectrum through DI water and DMP solution. The DMP has obvious UV absorption below 240 nm, while ozone absorbs UV light of short wavelengths in 200–300 nm with the maximum absorption coefficient at 253.7 nm. As shown in Fig. 4a and c, the UV lamp energy emitted was absorbed by DMP solution in UV-C region, while there was no absorption in UV-A and UV-B regions.

The results of Fig. 3 for the photolysis of DMP were obtained using the low-pressure UV lamp, yielding \( \eta_{\text{DMP}} \) of 2%. The corresponding removal efficiency (or mineralization efficiency) of TOC (\( \eta_{\text{TOC}} = (C_{\text{TOC}} - C_{\text{TOC}})/C_{\text{TOC}} \)) was relatively low. Thus, DMP is hard to be destructed via UV irradiation only. As UV irradiation was combined with HG-OZ system (HG-UV-OZ process), the decomposition of DMP reached 85% in 30 min reaction time (\( m_{\text{A,in}} = 600 \text{ mg L}^{-1} \)) and 99% in 60 min (\( m_{\text{A,in}} = 1200 \text{ mg L}^{-1} \)).

In UV-OZ process for the decomposition of DMP, molecules absorbing UV irradiation of UV-C region include:

1. molecular ozone – enhancing the decomposition of ozone in water and generating highly reactive hydroxyl radicals,
2. DMP molecule – absorbing UV-C via the aromatic ring,
3. intermediates from the decomposition of ozone – absorbing UV and reacting with ozone and hydroxyl radicals to yield other by-products (such as phenols–quinines–aromatic acids, short-chain aliphatic acids and short-chain aliphatic aldehydes).

The mechanism of UV-OZ may be described as follows. In aqueous phase, the dissolved ozone absorbs UV radiation to produce hydrogen peroxide (\( \text{H}_2\text{O}_2 \)) via the following reactions [17]:

\[
\text{O}_3 + \text{H}_2\text{O} \xrightarrow{h/\text{UV}} \text{O}_2 + \text{H}_2\text{O}_2 \quad (3)
\]

\[
\text{O}_3 + \text{H}_2\text{O} \xrightarrow{h/\text{UV}} \text{O}_2 + 2\text{OH}^* \quad (4)
\]

The second step of UV-OZ process is to form hydroxyl radical. The \( \text{H}_2\text{O}_2 \) may be decomposed further to form hydroxyl radical via photolysis or ozonation as follows.

Photolysis:

\[
\text{H}_2\text{O}_2 \xrightarrow{h/\text{UV}} 2\text{OH}^* \quad (5)
\]

Ozone enhanced decomposition:

\[
\text{H}_2\text{O}_2 \leftrightarrow \text{HO}_2^- + \text{H}^+ \quad (6)
\]

\[
\text{HO}_2^- + \text{O}_3 \rightarrow \text{HO}_2^* + \text{O}_3^- \quad (7)
\]

\[
\text{O}_3^- + \text{H}^+ \rightarrow \text{HO}_2^* \rightarrow \text{OH}^* + \text{O}_2. \quad (8)
\]

Comparing the reactions in the above pathway, the reaction rate of \( \text{H}_2\text{O}_2 \) via photolysis is quite slow (Eq. (5)). Therefore, the decomposition of \( \text{H}_2\text{O}_2 \) via ozone to form hydroxyl radical is the main pathway (Eqs. (6)–(8)) in this experiment.

The initial pH value of sample solution was about 7.5. For the HG and HG-UV processes, the pH values were remained at about 7.5 because of nearly no decomposition of DMP. Nevertheless, the pH value of DMP solution decreased during the ozonation because of the formation of intermediates such as aromatic acids and short-chain aliphatic acids. In the HG-OZ process, the pH value decreased from 7.5 to 6.0 and 3.6 in 5 and 15 min, respectively. As for the HG-UV-OZ process, the pH value changed from 7.5 to 5.9 and 6.2 in 5 and 15 min, respectively. At later period of HG-UV-OZ, the pH value of solution increased slightly because of further decomposition of organic acids.

3.3. Effects of \( m_{\text{A,in}} \) and \( N_r \) on the ozonation of DMP

The decomposition of DMP via the HG-OZ process at various \( N_r \) of HGPRP is shown in Fig. 5. The conditions of the process were: \( C_{\text{DMP}} = 100 \text{ mg L}^{-1} \), \( Q_r = 1 \text{ L min}^{-1} \), \( Q_{\text{A}} = 462 \text{ mL min}^{-1} \), \( V_l = 1 \text{ L} \) and \( C_{\text{AG,in}} = 25 \text{ mg L}^{-1} \). Five rotating speeds 600, 900, 1200, 1600 and 1800 rpm were tested to evaluate the effect of \( N_r \) on CDP. The corresponding gravitational forces resulted were 23.76, 53.48, 95.07, 148.55 and 296.31 m s⁻², respectively. The result showed that under the experimental conditions, the values of \( \eta_{\text{DMP}} \) increased significantly with \( N_r \). As \( N_r \) increased from 600 to 1800 rpm, the \( \eta_{\text{DMP}} \) at \( m_{\text{AG,in}} = 1200 \text{ mg L}^{-1} \) increased from 83 promote to 98%. High rotating speed can enhance the ozone gas–liquid mass transfer, thus resulting in a high \( \eta_{\text{DMP}} \). Lin and Liu [13] also noted that an increase of \( N_r \) for a HG-OZ process increases the decolorization of RB19 because the centrifugal force increases the interfacial area while decreases the resistance of ozone mass transfer via decreasing the thickness of liquid film and size of droplets. The results of Fig. 5 further showed that as \( N_r \) increased higher than 1200 rpm, the value of \( \eta_{\text{DMP}} \) only increases slightly, indicating the minimiz-
tion of mass transfer resistance. As mentioned previously in Section 3.1, a higher \(N_t\) results in a shorter HRT, implying that the time for ozone gas–liquid mass transfer becomes shorter. Although a high \(N_t\) can decrease the liquid-film thickness and droplet size to improve the ozone gas–liquid mass transfer rate, however, the short contact time decreases the amount of ozone transferred.

3.4. Decomposition of DMP via catalytic ozonation

Fig. 6 presents the decomposition of DMP with and without Pt/\(\gamma\)-Al\(_2\)O\(_3\) (Dash 220N) during the ozonation in HGRPB. For the case without catalyst, glass beads were used. The value of \(\eta_{\text{DMP}}\) for HG-Pt-OZ process reaches 92% at \(m_{\text{A,in}} = 600\ \text{mg L}^{-1}\) (in 30 min). The decomposition of DMP is nearly completed at \(m_{\text{A,in}} = 1200\ \text{mg L}^{-1}\) (in 60 min). A comparison with HG-OZ indicates that DMP is decomposed rapidly because Dash 220N promotes the dissolved O\(_3\) to form highly reactive OH\(^{•}\) radicals. The Pt on the Al\(_2\)O\(_3\) surface is effective for the decomposition of O\(_3\) in gaseous as well as aqueous phase [18,19]. The mechanism of the heterogeneous catalytic ozonation in water is shown in Fig. 7 which was explained by Legube and Vel Leitner [4]. Based on the mechanism, the Pt on the Dash 220N surface assists the generation of OH\(^{•}\) radicals. The organic molecules (DMP and intermediates) are adsorbed on the Dash 220N surface and oxidized via an electron-transfer reaction. Then the organic radical species are desorbed from the Dash 220N and oxidized by OH\(^{•}\) radical or O\(_3\).

Combination of HG-Pt-OZ with UV irradiation (HG-UV-Pt-OZ) can further enhance the decomposition of DMP. As shown in Fig. 6, after 10 min oxidation (\(m_{\text{A,in}} = 200\ \text{mg L}^{-1}\)), \(\eta_{\text{DMP}}\) of HG-UV-Pt-OZ is 65%, while that of HG-Pt-OZ without UV irradiation is 48%. The value \(\eta_{\text{DMP}}\) of HG-UV-Pt-OZ and HG-Pt-OZ at 30 min (\(m_{\text{A,in}} = 600\ \text{mg L}^{-1}\)) are 94% and 92%. Both of these two catalytic ozonation processes (with Pt/\(\gamma\)-Al\(_2\)O\(_3\) or UV radiation) can completely decompose the DMP at 50 min (\(m_{\text{A,in}} = 1000\ \text{mg L}^{-1}\)).

3.5. Decomposition of TOC via various Higee processes

Fig. 8 compares the values of \(\eta_{\text{TOC}}\) of the HG related processes. After 1 h reaction, the HG-UV-OZ and HG-Pt-OZ have about the same decomposition efficiency of TOC (\(\eta_{\text{TOC}}\)) of 56% and 57%, respectively. The result showed that both the UV irradiation and Pt catalyst can enhance the decomposition of dissolved ozone to form OH\(^{•}\) radicals, assisting the decomposition of TOC. Although ozone was fed and transferred in the HGRPB, the organics in solution were oxidized in HGRPB as well as in photolysis reactor with residual ozone. The transferred ozone is adsorbed on the Dash 220N surface and forms OH\(^{•}\) radical to react with organics during the course of catalytic ozonation. The radicals in turn enhance the decomposition reactions of DMP and intermediates.

To further promote the decomposition ability of organic in the HG-UV-OZ and HG-Pt-OZ processes, the UV irradiation is combined with Pt/\(\gamma\)-Al\(_2\)O\(_3\) catalyst in HG-OZ forming a novel HG-UV-Pt-OZ process.
process. Experiment result showed that, under the same ozone dosage, the HG-UV-Pt-OZ process has the highest $\eta_{TOC}$ of 68%, comparing with other processes examined in this study.

A simplified scheme of the decomposition pathways of DMP for the catalytic ozonation with UV is shown in Fig. 9. Both the DMP and intermediates are attacked by $O_3$ and OH$^*$ radical. The initial attack of oxidants on DMP during ozonation is via the electrophilic addition to form aromatic acids. Then the aromatic rings are broken down by $O_3$ and OH$^*$ radical to form short-chain organic acids such as oxalic acid, glyoxalic acid and formic acid. These short-chain organic acids are further decomposed to CO$_2$ and H$_2$O eventually.

4. Conclusions

The decomposition of DMP via HG-OZ, HG-UV-OZ, HG-Pt-OZ and HG-UV-Pt-OZ is investigated in this paper. Several conclusions can be drawn as follows.

1. Packing Pt/$\gamma$-Al$_2$O$_3$ catalyst (Dash 220N) in HG-OZ process can effectively enhance the decomposition of DMP under the experimental conditions of this study. The $\eta_{TOC}$ of DMP increases from 38% of HG-OZ to 57% of HG-Pt-OZ because of the formation of OH$^*$ radicals.

2. The decomposition rate of DMP is dependent on the rotating speed $N_r$ of packing bed. The thickness of liquid film and droplet size decrease with highly centrifugal force. As $N_r$ is higher than 1200 rpm, the effect of $N_r$ is greatly reduced.

3. The use of UV irradiation in HG-Pt-OZ process significantly enhances both the decomposition and mineralization rate of DMP. The HG-UV-OZ process mainly absorbs UV-C ($\lambda = 200–280$ nm) energy to decompose ozone to form high active radicals. The photolysis of DMP via UV irradiation (HG-UV) only is negligible.

4. The combination of UV radiation in HGRPB system can enhance the decomposition of dissolved ozone to form OH$^*$ radicals, assisting the decomposition of TOC.

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References


