Comparison of electrospray ionization, atmospheric pressure chemical ionization and atmospheric pressure photoionization for determining estrogenic chemicals in water by liquid chromatography tandem mass spectrometry with chemical derivatizations

Guang-Wen Lien, Chia-Yang Chen *, Gen-Shuh Wang

Institute of Environmental Health, College of Public Health, National Taiwan University, Taipei 10055, Taiwan

**Abstract**

This study compared the sensitivities and matrix effects of four ionization modes and four reversed-phase liquid chromatographic (LC) systems on analyzing estrone (E1), 17β-estradiol (E2), estriol (E3), 17α-ethinylestradiol (EE2), 4-nonylphenol (NP), 4-tert-octylphenol (OP), bisphenol A (BPA) and their derivatives of dansyl chloride or pentafluorobenzyl bromide (PFBBr) in water matrices using a triple-quadrupole mass spectrometer with selected reaction monitoring (SRM). The four probes were electrospray ionization (ESI), atmospheric pressure chemical ionization (APCI), atmospheric pressure photoionization (APPI) and APCI/APPI; the four LC systems were ultra-performance liquid chromatography (UPLC) with or without post-column split, a mixed-mode column and two-dimensional LC (2D-LC). Dansylated compounds with ESI at UPLC condition had the most intense signals and less matrix effects of the various combinations of ionization and LC systems. The on-column limits of detection (LODs) of dansylated estrogens by SRM were 0.05–0.20 pg, and the LODs in sewage treatment plant effluent and in river water were 0.23–0.52 and 0.56–0.91 ng/L, respectively. The LODs using selected ion monitoring (SIM) reached low ng/L levels in real samples and measured concentrations were comparable with those of SRM.

© 2008 Elsevier B.V. All rights reserved.

**1. Introduction**

Feminizing contaminants of steroid estrogens, detergent degraders and plasticizers have caused a worldwide concern. They may influence the ecosystem at trace levels and affect human health through their contamination of drinking water. Natural estrogens 17β-estradiol (E2) and its synthetic analogue 17α-ethinylestradiol (EE2), an ingredient in oral contraceptives, are the most estrogenic. Moreover, their major metabolites, estrone (E1) and estriol (E3), are still bioactive. These steroid estrogens enter the water environment via the urine of humans and animals in the form of hydrophilic glucuronide and sulfate conjugates [1], which are biologically inactivated [2]. However, they are likely to be deconjugated in sewage treatment systems and converted to estrogenically active free forms [3]. 4-nonylphenol (NP), 4-tert-octylphenol (OP) and bisphenol A (BPA), which are all xenoestrogens, can affect normal endocrine functions. Although they are less potent, they are usually found in much higher concentrations in water (ng/L–μg/L) [3–6]. These xenoestrogens are released into the water environment from daily usage of non-ionic surfactants and plasticizers.

Atmospheric pressure photoionization (APPI) is an emerging source, which is capable of ionizing nonpolar compounds and is possibly less susceptible to matrix effects. In addition, dual-source ionization (e.g. atmospheric pressure chemical ionization (APCI)/APPI combo in this study) expands the range of compounds that can be simultaneously analyzed. Although most studies determined feminizing chemicals with electrospray ionization (ESI) coupled with LC/MS(/MS)[7–9], the suitability of APCI and APPI deserve further exploration.

Matrix effect, which co-eluting components from the matrix or the mobile phase may enhance or suppress signals, is an important issue in using LC/MS/MS. Selective extraction, additional clean-up, efficient LC separation or change of mobile phase compositions may reduce matrix effects [10]. Furthermore, while the use of suitable internal standards (e.g. isotope-labeled chemicals) may correct signal irreproducibility, this approach will not be able to overcome the loss in sensitivity caused by matrix effects. Some studies utilized direct online extraction or post-column split to minimize matrix effects and simplify the sample preparation. A novel column developed on September 2006 combines both size exclusion and
reverse-phase chemistry to separate small molecules from complex matrix [11]; to the best of our knowledge, one study has analyzed drugs in bovine serum using the mixed-mode column [11]. A restricted access material (RAM) pre-column, with a similar separation mechanism to the mixed-mode column, has also been applied on analyzing food, biological and environmental samples [12,13]. However, the RAM pre-column is an alkyl-diol silica (ADS) column and provides little chromatographic separation for small molecules; it requires a column switch to connect it with an additional analytical column for chromatographic separation (a two-dimensional LC, 2D-LC). In addition, a post-column split delivers only a portion of LC flow into the MS, which may substantially decrease matrix effects, especially when a flow rate into ESI interface was decreased to nanoflow of 0.1 μL/min [14,15]. This nanosplit requires special nanospray probes, which is not amenable to a conventional ESI interface, whose flow rates can be as low as 20–50 μL/min. Reports on the mixed-mode column, 2D-LC and post-column split are very limited in environmental analysis and so little is known about their ability to reduce matrix effects.

Recently, there has been an increase in the number of studies using ultra-performance liquid chromatography (UPLC) combined with MS/MS. UPLC takes advantage of smaller packing particles (<2.0 μm) that enable high flow rates for fast chromatography without sacrificing separation efficiency, and signal-to-noise (S/N) ratios of analytes are increased because of sharp peaks. However, to best of our knowledge none have used UPLC/MS/MS to study estrogenic compounds in water.

Steroid estrogens and phenolic xenoestrogens are weak acids and their ionization on ESI and APCI are not very efficient compared with other more polar chemicals. Chemical derivatization can add on moieties improving ionization and enhance signals. For example, dansyl chloride or pentafluorobenzyl bromide (PFBBr) can react with phenolic groups, significantly improving sensitivity [16–18]. By adding the dansyl moiety with ESI interface, signal intensity may be increased as much as three orders of magnitude [16, 19, 20]. This technique has been also found to improve the sensitivity in APIE interface when used to measure steroid estrogens [21]. To date no dansyl derivatives have been analyzed with APPI interface. PFBBr derivatives can capture soft electrons in APCI, resulting in unstable metastatic ions, and cause subsequent dissociation to generate negative ions through the loss of pentafluorobenzyl radical (electron-capture atmospheric pressure negative ionization, EC-APNI). For estrone, the use of PFBBr derivatives in EC-APNI can enhance efficiency of ionization as much as 25 times that of APCI alone [18]. This method has been also used in APPI with a high toluene dopant flow rate (e.g. 200 μL/min or higher) and was found to be able to detect as little as 0.17 pg of 2,4-dinitrophenol [22], whereas for PFBBr-derivatized estrone, signal enhancement (1.4–9.8 times) was less than that using EC-APNI [18]. Our group previously reported that dansylated estrogens with ESI interface provided better signal intensities than that PFBBr derivatives with EC-APNI, but obvious signal suppression was encountered with ESI when analyzing complex matrices such as river water and effluents from sewage treatment plants [23].

In this study, we investigated signal intensity and matrix effects on various chromatographic systems (UPLC with or without flow split, mixed-mode column, 2D-LC) and several ionization modes (ESI+, ESI−, APCI+, APCI−, APPI+, APPI−, APCI/APPI+, APCI/APPI−) for both estrogenic compounds and their derivatives of dansyl chloride and PFBBR. In addition, the study is unique in that it first optimized the operation conditions specific for each ionization methods, including those for LC columns, mobile phase flow rates and compositions. Previous studies usually compared the performance of different ionization sources under only one analytical column kept at a constant solvent flow rate, isocratic chromatography, the same injection volume, or flow injection analysis alone [24,25]. However, the conclusions based on the results using non-optimized parameters of various ionization methods could be controversial. The main purpose of the study was to find out the best combination of a chromatographic system and an ionization method with satisfactory sensitivity using low volumes of water samples or single quadrupole MS. The final method was validated using river water and effluents from a sewage treatment plant (STP).

2. Experimental

2.1. Chemicals and reagents

 Estrone, 17β-estradiol, estriol, 17α-ethinylestradiol, 4-tert-octylyphenol, bisphenol A, and bisphenol A-d16 (as a recovery standard) were obtained from Sigma/Aldrich (Saint Louis, MO, USA; purity > 98%). The technical mixture of nonylphenol was supplied by Riedel-de Haën (Seelze, Germany; purity > 94%). 2,4,16,16-2D4-estrone, 2,4,16,16-2D4-β-estradiol, 2,4,17-2D3-16α-hydroxy-17β-estradiol, 2,4,16,16-2D4-17α-ethinylestradiol and 4-n-Octyl-d17-pheno were bought from C/D/N Isotopes (Pointe-Claire, Quebec, Canada; purity > 98%). Bisphenol A-13C12 was purchased from Cambridge Isotope Laboratories (Andover, MA, USA; purity > 98%). Dalsyl chloride (5-(dimethylamino) naphthalene-1-sulfonyl chloride, ~95% purity), pentafluorobenzyl bromide (PFBBBr, purity > 99%), 4-methylmorpholine (purity > 99.5%), sodium hydrogen carbonate, and potassium hydroxide were purchased from Sigma/Aldrich. Milli-Q water was obtained from a Millipore water purification system (Milford, MA, USA). Formic acid (purity > 88%) and formaldehyde (purity > 37%) were provided by J.T. Baker (Phillipsburg, NJ, USA). Solvents, including methanol, acetone, n-heptane, acetonitrile and toluene, were all HPLC grade from J.T. Baker.

2.2. Extraction

The procedure used to extract estrogenic compounds from the water sample has been previously described [9]. Briefly, water samples were spiked with internal standards and then filtered through 90-mm PVDF membranes (pore size 0.45 μm) to remove suspended solids before extraction. Extraction was performed using 50-mm Bakerbond PolarPlus C18 Speedisks (J.T. Baker), followed by a cleanup using 40% methanol/60% Milli-Q water (v/v). The disks were dried for 10 min under a vacuum of about ~25 kPa. Analytes were eluted with three portions of 5-mL 50% methanol/50% dichloromethane (v/v). The eluates were filtered through 25-mm PTFE syringe filters (pore size 0.2 μm) and concentrated to dryness at 45 °C by a SpeedVac concentrator (Thermo Savant SPD 1010, Holybrook, NY, USA). The residues were re-dissolved by spiking recovery standard and then reacted with dansyl chloride reagents.

2.3. Derivatization

2.3.1. Dansyl chloride derivatization

The procedure used to derive dansyl chloride was based on EE2 derivatization method used by Penzes and Oertel [26] and Shou et al. [19]. Briefly, 0.9 mL of 100 μg/mL analytes in acetone was vortexed for 1.0 min with 0.1 mL of 1 mg/mL dansyl chloride in dry acetone followed by mixing with 0.01 mL of 0.1N sodium hydroxide for 1.0 min. The mixture was kept at 50 °C for 30 min. 5 mL of n-heptane was added to the mixture which was then shaken for 3 min. It was centrifuged at 3000 rpm for 10 min and refrigerated at ~20 °C. Once it had separated into two layers, the organic layer was collected and filtered through 0.20-μm PTFE into another glass tube. The aqueous layer was discarded. The organic layer was evaporated to dryness at 45 °C by a SpeedVac concentrator. The residue was re-dissolved
with 0.9 mL of methanol to optimize the parameters of operations on the MS. However, the above protocol cannot be directly applied to the derivatization of water samples and needs another protocol that was modified from Anari et al. [16] and Nelson et al. [21]. 100 μL of 0.9 ng/μL (for ESI analysis) and 250 μL of 0.36 ng/μL (for APPI, APCI and APCI/APPI analysis), both in acetone, were vortexed with 250- and 625-μL sodium bicarbonate buffer (10 mM, pH adjusted with NaOHaq) to 10.5, respectively. To these standards were added 250- and 625-μL acetone, respectively. They were then incubated at 60°C for 3.0 min and evaporated to dryness in a SpeedVac concentrator. The residues were reconstituted with 100 and 250 μL of methanol, respectively. Then, 4 μL (for ESI mode) and 10 μL (for APPI, APCI and APCI/APPI mode) were injected into LC/MS/MS for comparative analysis.

2.3.2. PFBBr derivatization

PFBBr was derivated based on a procedure reported by Singh et al. [18]. We vortexed 250-μL mixture standards of native (0.36 ng/μL) analytes in methanol with 250 μL of potassium hydroxide in anhydrous ethanol (8:1000; w/v). We then added 250 μL of 5% PFBBr in acetonitrile. The mixture was baked at 60°C for 30 min and then evaporated to dryness in a SpeedVac concentrator. The residue was reconstituted with 250 μL of methanol, and 10 μL was injected into LC/MS/MS for comparative analysis.

2.4. LC systems and analytical columns

2.4.1. The UPLC with or without post-column split

A Waters BEH C18 column (2.1 mm × 100 mm, 1.7 μm) was used for ESI and APPI at a flow rate of 0.5 mL/min, while dansylated analytes in APPI were set at 0.2 mL/min for better signal intensities. A Sepax GP-C18 column (3.0 mm × 100 mm, 1.7 μm) was used for APPI and APCI/APPI at a flow rate of 1.0 mL/min. Post-column split (split ratio = 1:5) was tested on ESI.

2.4.2. The mixed-mode column

A Shodex ODP 2 HP-2D (2.0 mm × 150 mm, 5 μm) was used for ESI and APPI at a flow rate of 0.2 mL/min, while dansylated analytes in APPI were set at 0.5 mL/min. A Shodex ODP 2 HP-4D column (4.6 mm × 150 mm, 5 μm) was used for APPI and APCI/APPI at a flow rate of 1.0 mL/min.

2.4.3. 2D-LC system with RAM pre-column

This system was composed of a VICI six-port switching valve (Valco Instruments Inc., Houston, TX, USA) and an extra isocratic pump (Jasco PU-980, Tokyo, Japan) (Fig. 1). The elution profiles of the RAM pre-column (Lichrosphere RP-4ADS (25 mm × 2 mm, 25 μm)) for effluent of sewage treatment plants and river water were monitored with a UV detector set at 280 nm. The injection volume was 50 μL of each extract, and the flow rate of mobile phase of Milli-Q water-acetonitrile (95:5, v/v) through RAM pre-column was set at 1 mL/min. The time required to elute major matrix components was less than 3 min (profile not shown). Based on this result, the valve was switched after 3 min to backflush the analytes into an analytical column. A Thermo Hypersil Gold column (2.1 mm × 50 mm, 1.9 μm, Bellefone, PA, USA) was used for ESI and APPI, which flow rate was 0.2 mL/min, while dansylated analytes in APPI was set at 0.5 mL/min. A Thermo BetaBasic C18 column (4.6 mm × 150 mm, 3 μm) was used for APPI and APCI/APPI set at a flow rate of 1.0 mL/min. All chromatographic separations including UPLC, mixed-mode column and 2D-LC were performed at 60°C (Table 1).

2.5. Instruments and parameters

The separation and detection were performed on a Waters Acquity UPLC system (Waters Corporation, Milford, MA, USA) coupled with a Waters Quattro Premier XE triple quadrupole mass spectrometer. The UPLC/MS/MS system was controlled by

---

**Table 1**

<table>
<thead>
<tr>
<th>Column type</th>
<th>Native (-)</th>
<th>PFBBr (+)</th>
<th>Dansyl chloride (+)</th>
</tr>
</thead>
<tbody>
<tr>
<td>BEH C18 (2.1 mm × 100 mm, 1.7 μm)</td>
<td>ESI</td>
<td>APCI</td>
<td>ESI</td>
</tr>
<tr>
<td>GP-C18 (3.0 mm × 100 mm, 1.7 μm)</td>
<td>APCI/([APPI)</td>
<td>APCI</td>
<td>APCI/([APPI)</td>
</tr>
<tr>
<td>Mixed-mode LC</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>ODP 2 HP-2D (2.0 mm × 150 mm, 5 μm)</td>
<td>ESI</td>
<td>APCI/([APPI)</td>
<td>ESI</td>
</tr>
<tr>
<td>ODP 2 HP-4D (4.6 mm × 150 mm, 5 μm)</td>
<td>ESI</td>
<td>APCI/([APPI)</td>
<td>ESI</td>
</tr>
<tr>
<td>2D-LC with RAM</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Hypersil Gold (2.1 mm × 50 mm, 1.9 μm)</td>
<td>ESI</td>
<td>APCI</td>
<td>ESI</td>
</tr>
<tr>
<td>BetaBasic C18 (4.6 mm × 150 mm, 3 μm)</td>
<td>ESI</td>
<td>APCI/([APPI)</td>
<td>ESI</td>
</tr>
</tbody>
</table>
Table 2
Major MS parameters for different analytes and ionization methods.

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Native</th>
<th>PFBBr</th>
<th>Dansyl chloride</th>
</tr>
</thead>
<tbody>
<tr>
<td>Source temperature (°C)</td>
<td>120</td>
<td>140</td>
<td>120</td>
</tr>
<tr>
<td>Desolvation temperature (°C)</td>
<td>400</td>
<td>600</td>
<td>400</td>
</tr>
<tr>
<td>Cone gas flow (L/h)</td>
<td>50</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>Desolvation gas flow (L/h)</td>
<td>900</td>
<td>75</td>
<td>150</td>
</tr>
<tr>
<td>Corona (μA)</td>
<td>400</td>
<td>600</td>
<td>600</td>
</tr>
<tr>
<td>Capillary (kV)</td>
<td>3</td>
<td>2</td>
<td>2.8</td>
</tr>
<tr>
<td>Repeller (kV)</td>
<td>2</td>
<td>2</td>
<td>3</td>
</tr>
</tbody>
</table>

MassLynx V4.1 with QuanLynx Application Manager and the data were acquired and processed using MassLynx V4.1. Instrumental parameters in various ionization methods were optimized to achieve maximal analyte signal intensities.

Changes of the desolvation gas (N₂) flow rate and source temperature did not produce significantly different signals in either ESI or APCI interface when flow rates of mobile phase were at either 0.2 or 0.5 mL/min, so these two parameters were kept the same for these two flow rates. Major parameters are summarized in Table 2. Extractor voltage was 3.0 V and RF lens voltage was 0 V. Collision gas was argon at 3 × 10⁻³ mbar. Ion energy 1 and 2 were set at 0.3 and 3, respectively. Both LM 1 and LM 2 resolution were set at 15. The multiplier voltage was set at 650 V. Ions were monitored by selected reaction monitoring (SRM) as shown in Table 3. Dansylated analytes produced intense precursor ions with the same for these two flow rates. Major parameters are summarized in Table 2. Extractor voltage was 3.0 V and RF lens voltage was 0 V. Collision gas was argon at 3 × 10⁻³ mbar. Ion energy 1 and 2 were set at 0.3 and 3, respectively. Both LM 1 and LM 2 resolution were set at 15. The multiplier voltage was set at 650 V. Ions were monitored by selected reaction monitoring (SRM) as shown in Table 3. Dansylated analytes produced intense precursor ions with m/z [M+233.8]+, and the collision-induced dissociation produced intense product ions with m/z 171+ and m/z 156+, corresponding to the 5-(dimethylamino)-naphthalene moiety and the loss of one methyl group from the m/z 171+, respectively [19,23]. PFBBr derivatives produced the same precursor ions as the underivatized ones with m/z [M-H]+ [23]. Several LC mobile phase compositions were tested to obtain good separation and peak shapes. Data points across the peak were no less than 20 to ensure the integration precision.

2.5.1. Dansyl derivatives

2.5.1.1. ESI (+). 10 mM formic acid (pH 2.9) (A) and acetonitrile (B) were used as the mobile phase. There were three LC conditions. (1) A GP-C₁₈ column had an initial gradient of 50% B, followed by a linear gradient to 85% B for 0.5 min, and then to 100% B in 1.7 min, at which point it was held at 100% B for 0.3 min before being returned to initial condition. The column was re-equilibrated for 2.0 min.

2.5.1.3. APCI (+) and APCI/APPI (+). 10 mM formic acid (A) and methanol (B) were used as the mobile phase. There were three LC conditions. (1) A GP-C₁₈ column had an initial gradient of 50% B, followed by a linear gradient to 85% B for 0.5 min, and then to 100% B in 1.7 min, at which point it was held at 100% B for 0.3 min before being returned to initial condition. The column was re-equilibrated for 2.0 min. (2) An ODP 2 HP-4D column had an initial gradient of 50% B, followed by a linear gradient to 85% B for 3.0 min, and then to 100% B in 3.0 min. It was held at 100% B for 2.0 min before being returned to initial condition. The column was re-equilibrated for 2.0 min. (3) A RAM coupled with a BetaBasic C₁₈ column had a gradient of 30% B for 3 min, followed by a linear gradient to 50% B in 1.0 min, then to 85% B in 2.0 min, and then to 100% B in 2.0 min. It was held at 100% B for 1.0 min before being returned to initial condition. The column was re-equilibrated for 2.0 min.

2.5.2. PFBBr derivatives

2.5.2.1. APCI (−). Water (A) and methanol (B) were used as the mobile phase. There were three LC conditions. (1) A GP-C₁₈ column had a gradient of 70% B for 0.2 min, followed by a linear gradient to 90% B in 0.8 min, and then to 100% B in 2.2 min. It was held at 100% B for 0.2 min before being returned to initial condition. The column was re-equilibrated for 1.0 min. (2) An ODP2 HP-4D column had a gradient of 70% B for 2 min, followed by a linear gradient to 95% B in 3.0 min. It was held at 95% B for 2.0 min before being returned to initial condition. The column was re-equilibrated for 2.0 min. (3) A RAM coupled with a BetaBasic C₁₈ column had a gradient of 30% B for 3 min, followed by a linear gradient to 50% B in 1.0 min, then to 85% B in 2.0 min, and then to 100% B in 2.0 min, at which point it was held at 100% B for 2.0 min before being returned to initial condition. The column was re-equilibrated for 2.0 min.

2.5.3. Native analytes

2.5.3.1. ESI (−). 10 mM 4-methylmorphline (pH 9.5) (A) and acetonitrile (B) were used as mobile phase. There were three LC conditions. (1) A BEH C₁₈ column with and without split had a gradient of 50% B for 0.2 min, followed by a linear gradient to 85% B in 0.8 min, and then to 100% B in 1.5 min, at which point it was held at 100% B for 0.7 min before being returned to initial condition. The column was re-equilibrated for 1.0 min. (2) An ODP 2 HP-2D column had a gradient of 50% B for 2.0 min, followed by a linear gradient to 85% B in 1.0 min, then to 90% B in 2.0 min, at which point it was held at 90% B for 1.5 min before being returned to initial condition. The column was re-equilibrated for 4.5 min. (3) A RAM coupled with a Thermo Hypersil Gold column had a gradient of 30% B for 3.0 min, followed by a linear gradient to 50% B in 1.0 min, then to 85% B in 2.0 min, and then to 100% B in 2.0 min before being returned to initial condition. The column was re-equilibrated for 2.0 min.

2.5.3.2. APCI (+). 10 mM formic acid (A) and acetonitrile (B) were used as the mobile phase. There were three LC conditions. (1) A BEH C₁₈ column had a gradient of 10% B for 0.2 min, followed by a linear gradient to 40% B in 0.8 min, then to 70% B in 1.7 min, and then to 95% B in 0.5 min. It was kept at 95% B for 0.6 min before being returned to the initial condition. The column was re-equilibrated for 1.7 min. (2) An ODP 2 HP-2D column had a gradient of 10% B for 2.0 min, followed by a linear gradient to 50% B in 3.0 min, and then to 70% B in 8.0 min, at which point it was held at 70% B for 2.0 min before being returned to initial condition. The column was re-equilibrated for 4.0 min. (3) A RAM coupled with a Thermo Hypersil Gold column had a gradient of 30% B for 3.0 min, followed by a linear gradient to 50% B in 1.0 min, then to 75% B in 3 min, and then to 100% B in 2 min. It was held at 100% B for 2.0 min before being returned to initial condition. The column was re-equilibrated for 2.0 min.
2.5.3.2. **APPI (−)**. Water (A) and methanol (B) were used as mobile phase. There were three LC conditions. (1) A BEH C18 column had a gradient of 30% B for 1.0 min, followed by a linear gradient to 50% B in 2.0 min, then 75% B in 4.0 min, and then 100% B in 2.0 min. It was held at 100% B for 2.0 min before being returned to initial condition. The column was re-equilibrated for 4.0 min. (2) An ODP 2 HP-2D column had a gradient of 50% B for 1.0 min, followed by a linear gradient to 75% B in 1.5 min, and then to 90% B in 5.0 min, where it was held at 90% B for 1 min before being returned to initial condition. The column was re-equilibrated for 4.0 min. (3) A RAM coupled with a Thermo Hypersil Gold column had a gradient of 30% B for 3.0 min, followed by a linear gradient to 50% B in 1.0 min, then to 75% B in 3.0 min, and then to 100% B in 2.0 min, at which point it was held at 100% B for 2.0 min before being returned to initial condition. The column was re-equilibrated for 2.0 min.

2.5.3.3. **APCI (−) and APCI/APPI (−)**. Water (A) and methanol (B) were used as mobile phase. There were three LC conditions. (1) A GP-C18 column had an initial gradient of 30% B, followed by a linear gradient to 50% B for 0.5 min, then to 75% B in 2.0 min, and then to 100% B in 0.5 min. It was held at 100% B for 0.6 min before being returned to initial conditions. The column was re-equilibrated for 4.0 min. (2) An ODP 2 HP-4D column had a gradient of 50% B for 2.0 min, followed by a linear gradient to 75% B in 1.5 min, and then to 90% B in 3.5 min, at which point it was held at 100% B for 1.0 min before being returned to initial condition. The column was re-equilibrated for 4.0 min. (3) A RAM coupled with a BetaBasic C18 column had a gradient of 30% B for 3.0 min, followed by a linear gradient to 75% B in 1.0 min, and then to 100% B in 4.0 min, where it was held at 100% B for 4.0 min before being returned to initial condition. The column was re-equilibrated for 3.0 min.

### Table 3

<table>
<thead>
<tr>
<th>Analyte</th>
<th>MW</th>
<th>[M+H]+</th>
<th>Native</th>
<th>PFBB</th>
<th>Dansyl chloride</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td>CE:CV</td>
<td>CE:CV</td>
<td>CE:CV</td>
<td></td>
</tr>
<tr>
<td>E1</td>
<td>270.4</td>
<td>269.1 &gt; 145.0</td>
<td>304.1 &gt; 171.1</td>
<td>34</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>E1-d4</td>
<td>274.4</td>
<td>273.4 &gt; 146.8</td>
<td>308.1 &gt; 171.0</td>
<td>34</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>E2</td>
<td>272.4</td>
<td>271.2 &gt; 183.0</td>
<td>308.1 &gt; 171.0</td>
<td>46</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>E2-d4</td>
<td>276.1</td>
<td>274.6 &gt; 147.0</td>
<td>310.1 &gt; 171.0</td>
<td>46</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>E3</td>
<td>288.4</td>
<td>287.2 &gt; 170.9</td>
<td>322.2 &gt; 171.1</td>
<td>34</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>E3-d3</td>
<td>291.4</td>
<td>290.6 &gt; 173.0</td>
<td>325.2 &gt; 171.0</td>
<td>34</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>EE2</td>
<td>296.4</td>
<td>295.2 &gt; 144.8</td>
<td>330.2 &gt; 171.1</td>
<td>34</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>EE2-d4</td>
<td>300.4</td>
<td>298.8 &gt; 147.0</td>
<td>334.2 &gt; 171.0</td>
<td>34</td>
<td>0.0001–1</td>
</tr>
<tr>
<td>NP</td>
<td>220.4</td>
<td>219.2 &gt; 133.0</td>
<td>345.2 &gt; 171.2</td>
<td>32</td>
<td>0.05–3</td>
</tr>
<tr>
<td>OP</td>
<td>206.3</td>
<td>205.0 &gt; 133.2</td>
<td>340.2 &gt; 171.1</td>
<td>34</td>
<td>0.1–2</td>
</tr>
<tr>
<td>4-n-Octyl-d17-phenol</td>
<td>223.4</td>
<td>222.1 &gt; 107.6</td>
<td>457.2 &gt; 171.0</td>
<td>34</td>
<td>0.05–3</td>
</tr>
<tr>
<td>BPA</td>
<td>228.3</td>
<td>227.2 &gt; 212.1</td>
<td>462.1 &gt; 171.2</td>
<td>38</td>
<td>0.05–3</td>
</tr>
<tr>
<td>BPA-d16 (RS)</td>
<td>240.2</td>
<td>238.9 &gt; 223.5</td>
<td>478.0 &gt; 171.0</td>
<td>38</td>
<td>0.05–3</td>
</tr>
</tbody>
</table>

Note: 4-n-Octyl-d17-phenol was used as internal standard of NP and OP.

Because estrogenic compounds are frequently observed in sewage or surface water, it is difficult to obtain a matrix without estrogenic compounds. Raw water from a drinking water treatment plant (WTP) in Taipei City, which only contains analytes at trace levels, was used as the matrix. Equal aliquots from extracts of one-liter samples were used for each method. Eluates were concentrated to dryness at 45 °C using a SpeedVac concentrator and were reconstituted by appropriate solvents, with or without the spiking of 90-ng native compounds for the following analyses: (1) native chemicals in ESI (−), APPI (−), APCI (−) and APCI/APPI (−); (2) dansyl derivatization in ESI (+), APPI (+), APCI (+) and APCI/APPI (+); (3) PFBB derivatization in EC-APNI (−). The best method was chosen for method validation based on signal intensities and matrix effects. The percentage matrix effect (%ME) was used to assess matrix effects: peak area of post-extraction spiking/peak area of standard × 100. Before calculation, the areas of samples without spiking were subtracted from the areas of samples.

### 2.7. Method validation

Two types of water, effluents and river water, were used for method validation. The effluents were sampled from a sewage treatment plant in Taipei. That plant is a secondary treatment facility with an activated sludge units. The samples were collected in...
January 2008 (pH 7.06, temperature = 21.0 °C, DO = 4.29 mg/L). The river water was taken from the Kee-Lung River in Taipei in April 2008 (pH 7.0, temperature = 22.4 °C, DO = 1.2 mg/L). Three solutions, 0.25 ng/μL of the four estrogens standards (20, 100, or 200 ng/μL), 1 ng/μL of the three xenoestrogens standards (50, 250, or 500 μg/μL) and 100 μL of 0.5 ng/μL internal standards, were spiked into 0.5-L water samples before extraction. Eluates from the SPE disk were concentrated to dryness at 45 °C using the SpeedVac concentrator and were re-dissolved with 200-μL anhydrous acetone containing 0.25 ng/μL recovery standard, and then reacted with dansyl chloride derivatization reagent. Four-microliter solution was injected into LC/MS/MS.

2.8. QA/QC, quantification and data analysis

All glassware was rinsed with acetone, n-heptane, dichloromethane and methanol before being used for experiments. A blank sample spiked with the internal standards was run with each batch of samples to check experimental contamination and provide background levels of the native analytes. A
calibration curve was built at each analysis. The linear range were $0.0001–1 \text{ ng/\mu L}$ for steroid estrogens (except E2 $0.001–1 \text{ ng/\mu L}$) and $0.05–3 \text{ ng/\mu L}$ for xenoestrogens (except OP $0.1–2 \text{ ng/\mu L}$) at weighted $(1/x)$. Isotope-dilution techniques were used to correct variations in sample preparation and instrumental performance; peak areas of dansylated analytes were normalized to their deuterium-labeled internal standard for quantification. The squares of correlation coefficients ($r^2$) were 0.995 or above for all calibration curves as shown in Table 3. A one-way analysis of variance (ANOVA) with Tukey's post hoc comparison was used to compare the signal intensities and matrix effects associated with the different methods. Student's $t$-test was used to evaluate the difference of quantitative results between SRM and selected ion monitoring (SIM) method. SAS 9.1 was used to perform statistical analysis.

3. Results and discussion

3.1. Effects of dopant, mobile phase flow rates and compositions on APPI sensitivity

We found that excess dopant flow may or may not improve analyte signals, and optimal dopant portions were compound-dependent. Different amount of the toluene dopant were tested on a
Waters BEH C18 column, ranging from 5% to 25% of a constant mobile phase flow rate at 500 μL/min. The mobile phase compositions for native and dansylated compounds were Milli-Q water/methanol and 10 mM formic acid/acetonitrile, respectively, which were the optimal combinations for their separation on the column. Based on peak areas, the optimal amount of dopant for native analytes was 5% of the mobile phase flow, 25 μL/min. The intensities of dansylated analytes were gradually enhanced as toluene was increased to 20% of the mobile phase flow (corresponding to 100 μL/min). All intensities were enhanced except for dansyl-NP and dansyl-OP, whose signals dropped significantly once the dopant exceeded 5%. Therefore, a five-percent dopant flow was used in this study. Robb et al. used acridine and 9-methylanthracene as model compounds and reported that the signals are close to plateaus when the dopant amount ranges from 5% to 10% of the mobile phase flows, which were at 50, 200, or 1000 μL/min [27] and signal intensities are slightly raised when a higher dopant percentage is given [27], finding similar to ours in native and dansylated steroid estrogens and dansylated BPA; however, in our study, signals of dansylated NP and OP dropped when the dopant portion was higher than 5%.

The suitable flow rate of mobile phase for APPI signals was compound-dependent, which the best flow rates were different between native and dansylated analytes. The mobile phase flow rates of 100, 200, 500 and 1000 μL/min were evaluated at a constant dopant of 5% mobile phase flows using an isocratic liquid chromatography; the mobile phase compositions of native and dansylated analytes were 20% Milli-Q water/80% methanol (v/v) and 12% 10-mM formic acid/88% acetonitrile (v/v), respectively. A Sepax GP-C18 column (3.0 mm × 100 mm, 3 μm) was used at a flow rate of 1.0 mL/min and a Waters BEH C18 column (2.1 mm × 100 mm, 1.7 μm) was performed at equal or less than 500 μL/min. The best sensitivities of native and dansylated analytes were obtained at flow rates of 200 and 500 μL/min, respectively. The signal intensities of native analytes at a flow rate of 200 μL/min were 1.6- to 2.8-fold, 1.0- to 1.6-fold, and 2.7- to 6.4-fold higher than those at flows of 100, 500, and 1000 μL/min, respectively. The signal intensities of dansylated analytes at a flow rate of 500 μL/min were 4.9- to 18-fold, 3.1- to 9.4-fold, and 2.2- to 6.0-fold higher than those at flows of 100, 200, and 1000 μL/min, respectively. Some studies have demonstrated that a low flow rate (e.g., ≤100 μL/min) may improve the ionization efficiency of APPI.

<table>
<thead>
<tr>
<th>Table</th>
<th>Matrix effect factor (%) obtained from raw water samples using post-extraction spiking (n = 4).</th>
</tr>
</thead>
<tbody>
<tr>
<td>E1</td>
<td></td>
</tr>
<tr>
<td>E2</td>
<td></td>
</tr>
<tr>
<td>E3</td>
<td></td>
</tr>
<tr>
<td>EE2</td>
<td></td>
</tr>
<tr>
<td>NP</td>
<td></td>
</tr>
<tr>
<td>OP</td>
<td></td>
</tr>
<tr>
<td>BPA</td>
<td></td>
</tr>
</tbody>
</table>

Note: 3.6 ng equivalent of each analytes was injected.

| Waters BEH C18 column, ranging from 5% to 25% of a constant mobile phase flow rate at 500 μL/min. The mobile phase compositions for native and dansylated compounds were Milli-Q water/methanol and 10 mM formic acid(acetonitrile, respectively, which were the optimal combinations for their separation on the column. Based on peak areas, the optimal amount of dopant for native analytes was 5% of the mobile phase flow, 25 μL/min. The intensities of dansylated analytes were gradually enhanced as toluene was increased to 20% of the mobile phase flow (corresponding to 100 μL/min). All intensities were enhanced except for dansyl-NP and dansyl-OP, whose signals dropped significantly once the dopant exceeded 5%. Therefore, a five-percent dopant flow was used in this study. Robb et al. used acridine and 9-methylanthracene as model compounds and reported that the signals are close to plateaus when the dopant amount ranges from 5% to 10% of the mobile phase flows, which were at 50, 200, or 1000 μL/min [27] and signal intensities are slightly raised when a higher dopant percentage is given [27], finding similar to ours in native and dansylated steroid estrogens and dansylated BPA; however, in our study, signals of dansylated NP and OP dropped when the dopant portion was higher than 5%. The suitable flow rate of mobile phase for APPI signals was compound-dependent, which the best flow rates were different between native and dansylated analytes. The mobile phase flow rates of 100, 200, 500 and 1000 μL/min were evaluated at a constant dopant of 5% mobile phase flows using an isocratic liquid chromatography; the mobile phase compositions of native and dansylated analytes were 20% Milli-Q water/80% methanol (v/v) and 12% 10-mM formic acid/88% acetonitrile (v/v), respectively. A Sepax GP-C18 column (3.0 mm × 100 mm, 3 μm) was used at a flow rate of 1.0 mL/min and a Waters BEH C18 column (2.1 mm × 100 mm, 1.7 μm) was performed at equal or less than 500 μL/min. The best sensitivities of native and dansylated analytes were obtained at flow rates of 200 and 500 μL/min, respectively. The signal intensities of native analytes at a flow rate of 200 μL/min were 1.6- to 2.8-fold, 1.0- to 1.6-fold, and 2.7- to 6.4-fold higher than those at flows of 100, 500, and 1000 μL/min, respectively. The signal intensities of dansylated analytes at a flow rate of 500 μL/min were 4.9- to 18-fold, 3.1- to 9.4-fold, and 2.2- to 6.0-fold higher than those at flows of 100, 200, and 1000 μL/min, respectively. Some studies have demonstrated that a low flow rate (e.g., ≤100 μL/min) may improve the ionization efficiency of APPI.
resulting from lower photoabsorption by solvent [27–29]. However, our data showed that a lower flow rate may not provide a better efficiency in APPI. To confirm these observations, we further utilized a BEH C18 column with a smaller I.D. (1.0 mm × 100 mm, 1.7 µm). Its optimal flow rate is 100 µL/min. We found that the signal response was even weaker than that used a larger column I.D. (2.1 mm × 100 mm, 1.7 µm). Solvent molecules (e.g. methanol) may be involved in the ionization process, especially in the dopant-assisted APPI, and the ions formed by proton transfer are less affected by high flow rates than those formed via charge exchange [25,28,30,31]. This would explain the reason that the optimal mobile phase flow rates on APPI efficiencies were compound-dependent.

Mobile phase composition is also critical to APPI sensitivity, and is also compound-dependent. For native analytes, a Milli-Q water-methanol combination gave better responses with better peak shapes than those using water-acetonitrile; however, for dansylated analytes, a composition of 10 mM formic acid-acetonitrile provided two to three times higher sensitivity than a composition of 10 mM formic acid-methanol. Cai et al. indicated that methanol has a lower photoabsorption cross-section than acetonitrile, and its dimmers can be ionized by a Kr Lamp (acetonitrile cannot) [25]; consequently, use of methanol would theoretically provide a better response, which was correct for native analytes in this study but was not the case for dansylated analytes. However, Cal et al. also proposed that the APPI mechanism is much more complex, and other factors, such as the ionization potential of analytes and the relative response, which was correct for native analytes in this study but was not the case for dansylated analytes. However, Cal et al. also proposed that the APPI mechanism is much more complex, and other factors, such as the ionization potential of analytes and the relative proton affinity of solvents and analytes, may affect APPI efficiencies as well [25].

3.2. Comparison of signal intensity between derivatized and underivatized analytes

The best combinations of LC systems and ion sources for native and dansylated analytes were 2D-LC and UPLC coupled with ESI mode, respectively. Signals of all native analytes in ESI (−) mode were better than those in APPI (−), APCl (−) and APCI/APPI (−) except for NP and OP; there was no significant difference in signal intensities between APCI/APPI dual mode and APCI or APPI alone (Fig. 2). Within the ESI mode, 2D-LC was superior to UPLC and mixed-mode column in terms of signal intensities, but UPLC outperformed both 2D-LC and mixed-mode columns when APPI (−), APCl (−) and APCI/APPI (−) were used. For PFBBr derivatives using EC-APNI (−), it was better to use 2D-LC than UPLC for the analysis of E1, E2, E3 and E2E2, but it was better to use UPLC rather than 2D-LC for the analysis of NP, OP and BPA. With regard to the signal intensities of the dansylated analytes, ESI (+) were much better than other ion sources (except for NP); the APCI/APPI dual mode produced similar signal intensities with those of APPI alone, but was much inferior to those of APCI alone. In addition, use of UPLC produced much stronger responses of dansylated analytes relative to other two LC systems in all ion sources (Fig. 2). Based on the results, the on-line cleanup of the mixed-mode column and the RAM pre-column did not give them a decisive advantage over UPLC on detecting the analytes in raw water samples, especially for dansylated derivatives. The mixed-mode column and 2D-LC, a polymer-based column and a coupled-column system, respectively, offered good peak shapes at widths <0.3 min (Fig. 4); nevertheless, UPLC provided sharp peaks at a width of about 0.06 min (except for NP, which was 0.14 min due to a mixture of isomers).

Derivatization with dansyl chloride and PFBBr significantly improved the detection sensitivity relative to underivatized analytes, and the dansylated analytes produced much better responses in four ionization methods than PFBBr derivatives in EC-APNI (Fig. 3). The trends of signal enhancement of derivatization among various ionization methods were similar, which were
independent on the LC systems (Fig. 5). For example, on UPLC column, the signal enhancements were 859–8460 times, 354–4030 times, 23–472 times, 21–344 times and 5–41 times in ESI, APPI, APCI, APCI/APPI and EC-APNI, respectively, relative to the native analytes. In other words, the order of responses of derivatized analytes versus underivatized ones in ionization methods was ESI > APPI > APCI ≒ APCI/APPI > EC-APNI. In addition, post-column split (a flow rate at 100 μL/min after split) in this study did not increase response in ESI mode (Fig. 3). In contrast, Kloepfer et al. reported that a lower flow rate (e.g. down to 20 μL/min) can dramatically increase signal intensities of some analytes, although some other analytes were unaffected [32].

3.3. Matrix effect

It was inconclusive to determine which LC systems and ion sources were least susceptible to matrix effects (i.e., higher values of matrix effect factors) when using the raw water as the matrix (Table 4), which is a source for drinking water and may be cleaner than usual surface water; however, we did observe that derivatized analytes were less prone to matrix effects than underivatized ones. For native analytes, we found no significant differences in matrix effect factors among the different LC systems with ESI mode; however, the matrix effect was lower on 2D-LC for APPI, and it was lower on the mixed-mode column for APCI and APCI/APPI. There were no significant differences in matrix effect factors among LC systems for dansylated analytes in all sources (except for APPI) and for PFBBr derivatives in EC-APNI.

Because the signal intensities of dansylated analytes using ESI and APPI were much more intense than others, we further investigated the matrix effects of these two sources on dansylated analytes using river water, which is a more complex matrix comparing with the raw water and may contain higher levels of analytes. The endogenous analytes in the river water may influence our determination on matrix effect factors; to avoid this potential problem, we spiked stable isotope-labeled analytes to the residues after extraction (the levels were equivalent to 80 ng/L in the water) before derivatization with dansyl chloride. The matrix effect factors of ESI using UPLC, UPLC with post-column split, mixed-mode column and 2D-LC were 17.7%–70.3%, 22.0%–57.3%, 40.2%–60.4% and 16.1%–62.1%, respectively; the factor values of APPI using UPLC, mixed-mode column and 2D-LC were 15.7%–46.7%, 25.9%–49.9% and 22.1%–63.2%, respectively. Therefore, the matrix effect of ESI and APPI were similar under the same LC conditions, and none of the four LC systems could significantly eliminate ion suppression. The cutoff of molecular mass for RAM pre-column is about 15 kDa; consequently, the RAM column could not eliminate the matrix effect caused by small molecules [32]. In a previous study, Kloepfer et al. reported that a post-column split (down to 20–100 μL/min) not only enhances sensitivity but also reduces ion suppression by 40–60% in wastewater samples [32]; However, the UPLC with post-column split did not reduce matrix effects significantly comparing with that without split in our study.

3.4. Method validation

Because dansyl derivatization under UPLC coupled with ESI provided the best performance based on the sensitivity and matrix effect, we validated this method. Good accuracy and precision were obtained for calibration curves. The intra-day and inter-day accuracy were within 2% and 9%, respectively; the inter-day variations (RSD%) ranged from 0.88% to 14.6% and intra-day responses were almost identical. The linearity of calibration curves of steroid estrogens (1000-fold to 10,000-fold) are similar or better than those of Qin et al. (0.2–200 ng/mL, 1000-fold) [20] and Vulliet et al. (0.05–20 ng/mL, 400-fold) [33].

No steroid estrogens were detected in any laboratory blank, but NP, OP and BPA were detected using dansyl derivatization; the background levels in the reagent blanks of Milli-Q water were 60–130 ng/L. The possible sources of NP, OP and BPA could be the speedisk for solid phase extraction (packed in a plastic disk vessel), the Milli-Q water itself (plastic materials in the water-purified system) or the derivatization process. For example, during the derivatization procedure, reagents (NaHCO3/NaOH buffer and

Fig. 6. Chromatograms of dansyl-E1 in effluents from a sewage treatment plant between: (a) SRM and (b) SIM mode.
ground levels in the water. In our study, RSD% of all spiked samples method accuracy and precision (four duplicates each level). The into the water samples from the same sources to evaluate the gens and 100, 500 and 1000 ng/L for xenoestrogens) were spiked and BPA in river water in Taiwan have also been observed by Ding et al. [34,35].

Three different concentrations (10, 50, 100 ng/L for steroid estrogens and 100, 500 and 1000 ng/L for xenoestrogens) were spiked into the water samples from the same sources to evaluate the method accuracy and precision (four duplicates each level). The reported concentrations of spiked samples did not deduct the background levels in the water. In our study, RSD% of all spiked samples were all smaller than 15.5% except for NP and OP spiked at 100 ng/L (and 500 ng/L for NP), which were close to the endogenous levels of the samples (Table 5). We found that the measured concentrations were very close to the spiked levels if the backgrounds (no spike) were deducted (Table 5).

Because of the existing backgrounds of NP, OP, and BPA, it is impractical to calculate their limits of detection (LODs); therefore, we only reported the LODs (S/N = 3:1) of E1, E2, E3 and EE2, which ranged between 0.23 and 0.52 ng/L for STP effluents and between 0.56 and 0.91 ng/L for river water (Table 5). Qin et al. also reported good method detection limits (MDL) of 0.038–0.13 ng/L for river water of 500 mL using dansyl chloride derivatization [20]. In addition, Vulliet et al. showed excellent LODs of 0.01–0.20 ng/L without chemical derivatization for 1-L groundwater, which may result from good recoveries and a large-volume injection (100 μL) [33]. Although tandem-MS methods (SRM) provide better sensitivity and selectivity than those of single-MS methods (SIM), most labs cannot afford tandem-MS instruments. However, this study, for example, was able to detect steroid estrogens in all real samples of STP effluents with SIM after dansyl derivatization, and the measured levels of either spiked or non-spiked samples were similar to those using SRM (Fig. 6). The LODs of steroid estrogens in effluents using SIM were 1.03–1.75 ng/L. The on-column detection limits of dansylated steroid estrogens with SRM and SIM were 0.05–0.20 and 0.44–1.48 pg, respectively.

4. Conclusions

In this study, we present a quantitative method for the analysis of seven estrogenic compounds with dansyl derivatization in both SIM and SRM. With the improvement in sensitivity using UPLC and chemical derivatization method, environmental levels of these chemicals can be determined using a single MS instead of the more expensive tandem-MS. The instrumental throughput was significantly increased, with a run in 3.2 min plus 1-min re-equilibrium time. We exhaustively investigated the performance of common ionization probes and different LC systems. ESI is usually reported to be more subject to ion suppression than APCI and APPI, but this was not the case in this study. Although a mixed-mode column or the RAM pre-column did not substantially reduce the matrix effects better than UPLC for the water matrices we tested, their use with other matrices, such as food or tissues, are worth further explorations.

Acknowledgement

This work is supported by the National Science Council, Taiwan (NSC 96–2314-B-002-101-MY2).

References